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Soil Organic Carbon in Forest and Other Land Use Types at Bengkulu City, Indonesia

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Abstract

Conversion of natural forest into agricultural land uses has decreased soil organic carbon (SOC) and increased carbon emission into the atmosphere, but proper management of agricultural land can sequester carbon from the atmosphere and increase the SOC. This study was conducted to estimate the SOC content and storage in a forest, agroforestry land, oil palm plantation, and agricultural experimental field and to analyze the correlation between the SOC and other soil characteristics at Bengkulu City, Indonesia. Soil were sampled from the following depths: 0–10 cm, 10–20 cm, and 20–30 cm. The biomass of litter and ground cover was also sampled. This study found that the forest had the highest average SOC content from the three depths, and 0–30 cm depth SOC storage, while the agroforestry system had the lowest of both SOC content and storage. The 0–10 cm depth had the highest SOC content and storage, while the 20–30 cm depth had the lowest of both variables. The SOC was positively correlated with litter biomass, field capacity, exchangeable potassium, cation exchange capacity, and negatively correlated with bulk density and exchangeable calcium, but not correlated with total nitrogen and available phosphorus. High litter biomass input is the key to the maintenance of high SOC.

Keywords: agroforestry, litter biomass, palm plantation, carbon sequestration

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Introduction

A natural forest has a higher capability of storing carbon both in the soil and above ground than most other land uses (Ravindranath & Ostwald, 2008). When a forest is converted to other land uses, the soil organic carbon (SOC) decreases while carbon emission into the atmosphere increases (Lal, 2020). As much as 1.74 Mg ha⁻¹ year⁻¹ of SOC may be lost when a forest is converted to farmland (Deng et al., 2016). In Sulawesi, the conversion of natural forest to oil palm plantation has reduced SOC up to 50% (van Straaten et al., 2015). In North Sumatera, the conversion of natural forest into agricultural land has reduced SOC up to 93% (Wasis, 2012). To prevent the SOC from being lost and released into the atmosphere, the most popular method is reducing deforestation and forest degradation. Annually, the tropical primary forest can sequester 813% of global carbon from anthropogenic emissions (Mackey et al., 2020). Preventing deforestation in developing countries is, however, difficult as the demand for land is increasing. In Indonesia, much of the natural forest has been converted into plantations, mostly oil palm plantations (Khasanah et al., 2015). Another method of preventing carbon loss into the atmosphere is improving soil management because the SOC is larger than other carbon pools in the carbon cycle (Deng et al., 2016). With proper management, soil can sequester a great amount of carbon from the atmosphere and store it. Therefore, during COP21 in Paris, France proposed the 4 per 1,000 initiative to boost the sequestration of carbon from better soil management (Minasny et al., 2017). Inputs of organic carbon for soil include crop residues, decomposed roots, and manure (Navarro-Pedreño et al., 2021). Therefore, the SOC in farmland can be increased by returning plant residues to the soil (Stella et al., 2019) or adding fertilizer, especially manure (Kadlec et al., 2012). The application of biochar can stabilize soil aggregate which in turn may increase the SOC storage (Li et al., 2021). However, the management of soil has no effect on the SOC below 30 cm depth (Franzluebbers, 2021).

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Increasing the SOC not only reduces carbon emission into the atmosphere but also improves soil fertility (FAO, 2017) and crop yield (Oldfield et al., 2019). Increasing the SOC can also increase CEC, water availability, and soil aggregation leading to reduced soil compaction (Murphy, 2015). Due to its important role in improving soil properties, the amount of SOC can be used as a soil sustainability index (Signor et al., 2018). Improving the SOC is, therefore, very important in the tropic regions such as Sumatera which has experienced forest conversion for several decades, including into oil palm plantation. Study of the SOC in different land uses and water characteristic has been done by Hermawan et al., (2020), while Farrasati et al., (2020) has studied the SOC in oil palm plantation and its relation with several chemical properties of soil. So far, no study has been done to correlate the SOC with both physical and chemical properties in forests and other land use types in mineral soil in Sumatera.

This study was done to estimate the SOC content and

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storage in secondary forest, oil palm plantation, agroforestry land, and agricultural experimental field at Bengkulu City, Indonesia, and to analyze the correlation between the SOC and other soil characteristics.

Methods

Study sites This research was conducted in four land use types, i.e., urban forest, agroforestry land, oil palm plantations, and agricultural experimental field at Bengkulu City, Bengkulu Province, Indonesia (Figure 1 and Figure 2) from July to September 2020. The urban forest (hereafter referred to as forest) is an old (>30 years old) secondary forest located at the University of Bengkulu. The oil palm plantation is approximately 8 years old located in Kandang Limun Village. The agricultural experimental field (hereafter referred to as agricultural field) is a field used by the University of Bengkulu students and faculty members to conduct agricultural experiments with seasonal plants. The agroforestry lands are home gardens consisting of trees, shrubs, and herbs, located in Surabaya Village. Unfortunately, we had no detailed information on their soil types.

Soil sampling Soil sampling for organic carbon is done following the Indonesian National Standard (SNI)-7724 procedure (BSN, 2011) with slight modifications as follows: In one plot, measuring 5×5 m, soil samples were taken from 5 points (4 cardinal points plus 1 in the center), at three depths (0–10 cm, 10–20 cm, and 20–30 cm), and then samples from 5 points were combined for each depth. Sampling was repeated 4 times in each land use type with a total of 12 composite soil samples.

The soil samples were air-dried, ground, and sieved using a 2 mm screen for texture analysis, and a 0.5 mm screen for other analysis. Soil organic carbon was analyzed using Walkley-Black Method and determined with a spectrophotometer using 561 nm wavelength and soil pH using pH meter with a ratio of soil and distilled water of 1:2.5

(Eviati & Sulaiman, 2009). Soil texture was determined with the hydrometer method (Agus et al., 2006a).

Sampling for the bulk density was done using rings in the plot. In each plot, five samples were taken at each depth, so there were 15 samples for each plot. Total samples for each land use type were 60. Sampling for field capacity was also done using rings, 3 samples at each depth, with a total of 36 samples for each land use type. More ring samples were taken than composite samples because each composite sample represented five points, while each ring sample represented only one point. Bulk density was determined using the core method (Agus, et al., 2006b), and field capacity using the pressure plate method (Sudirman et al., 2006).

For other soil properties, i.e., total N, available P, exchangeable K, exchangeable Ca, CEC samplings were done only for 0–10 cm depth, with a total of 16 samples for four land use types. Total soil nitrogen was determined with the Kjeldahl method, phosphorus with Bray I method, potassium with 1N NH4O Acetate extraction before was analyzed using a flame photometer, calcium with EDTA titration method, CEC using 1N NH4OAcetate extraction (Eviati & Sulaiman, 2009). The soil samples were processed and analyzed at the Soil Science Laboratory, Faculty of Agriculture, The University of Bengkulu.

Litter and groundcover biomass sampling In every plot, ground cover and litter were sampled from the same location as the soil sampling, using a frame sizing 1×1 m. All the ground cover within the frame was cut and weighed. All the litter was collected and weighed. In one plot, 4 samples of ground cover and 4 samples of litter were taken, with a total of 64 samples of ground cover and 64 samples of litter for four land use types. Then, from each frame, 300 g of the ground cover and 300 g of the litter was put into separate paper bags to be oven-dried until the weight was constant. Then, the oven-dried weight was recorded to determine the biomass.



Figure 1 Study sites in Bengkulu City, Indonesia: urban forest, oil palm plantation, agroforestry system, and agricultural field.

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Figure 2 Land use types in the study sites: forest (A), agroforestry (B), oil palm plantation (C), and agricultural experimental field (D).

SOC storage determination Based on the SOC content determined using a spectrophotometer, the SOC storage for certain depth was calculated using the formula as shown in Equation [1] (BSN, 2011).

$$Cst = Sd \times \rho \times SOC \times 100$$

note: Cst = soil organic carbon storage (Mg ha⁻¹); Sd = depth of soil (cm); ρ = bulk density (g cm⁻³); SOC = soil organic carbon concentration.

Statistical analysis The effect of land use types and soil depth on soil characteristics was determined with analyses of variance using Microsoft Excel. When a significant difference was found, then the Least Significant Difference (LSD) test was calculated manually (Steel & Torie, 1980). Correlation between the SOC content and other soil characteristics was also analyzed.

Results and Discussion

SOC content and storage in different land use types and soil depths The SOC content and storage were significantly different among land use types and depth (Table 1 and Table 2; Figure 3A and Figure 3B). The forest had the highest average SOC content from 3 depths (M = 4.97%, SD = 0.81%), and 0–30 cm depth SOC storage (M = 157.08 Mg ha⁻¹, SD = 14.79 Mg ha⁻¹), while the agroforestry land had the lowest (M = 1.98%, SD = 0.59%, and M = 70.81 Mg ha⁻¹, SD = 16.85 Mg ha⁻¹, respectively (Table 1).

These results confirm other studies' reports that the forest has relatively higher SOC than other land use types. In the highland of Mexico, the SOC in the 15 cm-depth forest soil was 40.31 Mg ha⁻¹, while the SOC in the same depth of arable land was only 5.9 Mg ha⁻¹ (Fusaro et al., 2019). A study in Tunisia (Jendoubi et al., 2019) found that forests had the SOC of 1.09% higher than that in crop lands (0.87%), grazing lands (0.74%), and field crops (0.70%). In Sulawesi, the SOC of oil palm plantations was 50% of that in natural forests (van Straaten et al., 2015). In six countries of Africa, the conversion of forest to cropland caused the loss of 6.7 Mg ha⁻¹ year⁻¹. The SOC in the top 30 cm layer for ten year period (Namirembe et al., 2020).

The higher SOC in the forest was presumably due to the significantly higher litter biomass ($M = 279.95 \text{ g m}^2$, $SD = 41.80 \text{ g m}^2$) than the other land use types (Table 1). The litter biomass and SOC content were strongly correlated positively (r = 0.84; p-value < 0.0001) (Figure 4).

Surprisingly, in this study, the lands utilizing agroforestry system had the lowest SOC, unlike the general notion that the agroforestry system has high SOC. In fact, implementing agroforestry is one of several methods to increase carbon sequestration (Lal, 2020). A meta-analysis found that the lands with agroforestry system in humid and sub-humid tropics can increase their SOC up to 21% compared to in monoculture system (Muchane, 2020). In Sulawesi, the SOC in the forest was 87 Mg ha⁻¹, and in cocoa agroforestry was 80 Mg ha⁻¹, but those numbers were not statistically different (Kessler et al., 2012). In Chiapas communities in

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Table 1 Soil organic carbon and biomass input in different land use types in the study sites

Land use types	Average SOC content (%)	0-30 cm-depth SOC storage (Mg ha ⁻¹)	Ground cover biomass (g m ⁻²)	Litter biomass (g m ⁻²)
Forest	$4.97 \pm 0.81a$	157.08 ± 14.79a	$8.24 \pm 7.21a$	279.95 ± 41.80a
Agricultural field	$3.28 \pm 1.42b$	$113.32 \pm 18.11b$	157.26 ± 105.36 b	$133.24 \pm 81.65b$
Oil palm plantation	$2.79 \pm 1.19b$	91.02 ± 5.93 bc	$211.72 \pm 63.78b$	$123.51 \pm 23.71b$
Agroforestry system	$1.98 \pm 0.59c$	$70.81 \pm 16.85c$	$41.43 \pm 30.04a$	$79.30 \pm 27.78b$

Note: Numbers are mean and standard deviation. SOC content: the average values from the three depths: 0-10 cm; 10-20 cm; 20-30 cm. Different letters following values in the same column indicate significant differences among the values (p-value < 0.05), determined with Fisher's LSD test.

Table 2 Soil organic carbon content and storage at different depths in the study sites

Soil depth (cm)	SOC content (%)	SOC storage (Mg ha ⁻¹)	
0-10	$4.16 \pm 1.41a$	44.60 ± 15.58a	
0-20	$3.30 \pm 1.21a$	$36.29 \pm 12.07b$	
20-30	$2.31 \pm 1.35b$	$27.17 \pm 13.74c$	

Note: Different letters following values in the same column indicate significant differences among the values (p-value < 0.05), determined with Fisher's LSD test.

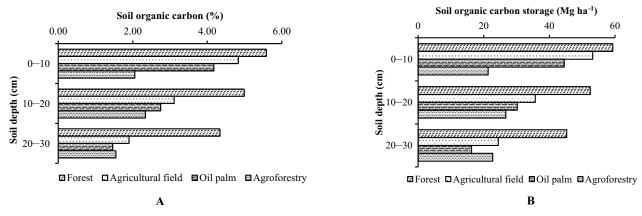


Figure 3 Soil organic carbon content (A) and soil organic storage (B) in different land use types and soil depths.

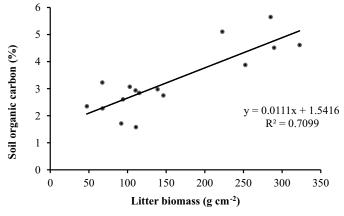


Figure 4 Correlation between SOC and litter biomass in the study sites.

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Mexico, the agroforestry system had higher soil carbon than the traditional maize cultivation and pastures without trees (Soto-Pinto et al., 2010).

The low SOC in the agroforestry system in the present study was presumably due to the lack of litter since the owners usually removed the litter (considered as garbage) from the agroforestry land, which was the home gardens located in the backyard of the houses. The agroforestry had the lowest litter, while the forest had the highest litter biomass (Table 1). The forest litter provided C input to the soil, thus maintaining high SOC.

In agricultural land, litter biomass from crop residues is one of the inputs of SOC (Navarro-Pedreño et al., 2021), so returning plant residues into the soil is the key to prevent SOC loss in agricultural land (Stella et al., 2019). In this study, the agricultural field which ranked second in the average SOC content and storage also ranked second in litter biomass (Table 1). The field was occasionally left fallow and grown with grass and sedges. Sometimes, it also received manure to improve the crop growth. A study in experimental plots near Prague, Czech Republic, showed that the input of organic matter, especially manure, reduced the SOC loss and increased carbon sequestration (Kadlec et al., 2012). In the Mediterranean cropping system, it was also found that crop residues and manure input were positively correlated with the SOC (Kong et al., 2005).

In the forest, the litter biomass was not contributed by the ground cover, because the forest had the lowest ground cover (Table 1), but by fallen tree leaves. The thick litter in the forest is decomposed into humus and enriches the soil with organic matter which is accumulated mostly in A horizon . It is not surprising, and therefore, the highest SOC content and storage in this study were found in the top layer.

In this study, the highest SOC content (M= 4.16 %, SD = 1.41%) and storage (M= 44.60 Mg ha⁻¹, SD = 15.58 Mg ha⁻¹) were found in 0–10 cm depth, the lowest (2.31% and 27.17 Mg ha⁻¹, respectively) were found in 20–30 cm depth (Table 2). In the agroforestry system of Chiapas communities in Mexico, the values in the same depth intervals also showed a decreasing pattern: 39.7, 23.2, and 18.1 Mg ha⁻¹ (Soto-Pinto et al., 2010). Soil organic carbon also decreased in deeper soil in the upper Yellow River of China (Guo et al., 2016), in the Lancangjiang River Basin, China (Zhou et al., 2020), and the tropical plain in Cameroon (Kome et al., 2021). However, despite having a lower concentration of carbon, deep soil has greater total SOC storage due to its greater soil mass (Gross

& Harrison, 2019).

Other soil properties and their correlation with SOC The forest soil had significantly higher cation exchange capacity (CEC) than the agroforestry system and oil palm plantation, and slightly, but not significant, higher than the agricultural field (Table 3). The SOC content was strongly correlated positively with soil CEC, r = 0.77, p-value < 0.001 (Figure 5A). A similar result was also reported from a study in Latosol soil in Mato Grosso State, Brazil, that after 10 years of no-till management, increase of SOC was positively correlated (r = 0.92) with CEC (Ramos, 2018). With higher CEC, the forest soil had more sites to bind cations such as K.

In this study, the forest soil also had the highest exchangeable K (Table 3), and the SOC was strongly correlated positively with exchangeable K, r=0.80, p-value < 0.001 (Figure 5B). Another study found that soil organic matter had more adsorption for K than mineral soil (Wang & Huang, 2001). In this study, the exchangeable Ca in forest soil was lower than the agroforestry land, but not significantly different from that in the agricultural field and oil palm plantation (Table 3), and SOC was correlated negatively with soil exchangeable Ca, r=-0.74, p-value < 0.01 (Figure 5E). In the upper Yellow River of China, Guo et al., (2016) who studied SOC and soil inorganic carbon (SIC) reported a positive correlation between SIC and soluble Ca²⁺, but did not report the correlation between SOC and Ca, presumably because SOC did not correlate with Ca.

The forest soil had the highest pH and pH was positively correlated with SOC, but the correlation was weak, r = 0.45, p-value < 0.01 (Figure 5C). In another study, SOC had no correlation with pH in inceptisols soil. However, in ultisols soils SOC was negatively correlated with pH, but the correlation was very weak, r = -0.141 (Farrasati et al., 2020).

The forest had higher P than the other land use types (Table 3), but SOC contents were not significantly correlated with P. The SOC is part of soil organic matter which generally contributes 2080% to the P soil, depending on soil type, and it prevents the P being fixed by Al and Fe and becomes unavailable to plant (Murphy, 2015).

In this study, the forest had higher N than other land use types, but the SOC had no correlation with N. Farrasati et al. (2020) did not find any correlation between SOC with N in ultisols soil, but they found that SOC was positively correlated with N in inceptisols, although the correlation was weak, with r = 0.392.

Table 3 Soil chemical properties in different land use types in the study sites

Landmastruss	рН	Ca	K	CEC	P	N. (0/)
Land use types		(cmol kg1)	(cmol kg1)	(cmol kg1)	(mg kg ¹)	N (%)
Forest	$5.29 \pm 0.28a$	$0.79 \pm 0.32a$	$0.45 \pm 0.05a$	$25.36 \pm 0.54a$	$11.04 \pm 1.50a$	$0.30 \pm 0.05a$
Agricultural field	$4.53 \pm 0.23c$	$1.24 \pm 0.30a$	$0.34 \pm 0.03b$	$22.86 \pm 0.96a$	$3.24 \pm 1.01c$	$0.21 \pm 0.10a$
Oil palm plantation	$4.62 \pm 0.09c$	$0.83 \pm 0.27a$	$0.38 \pm 0.04b$	15.21 ± 3.65 b	7.40 ± 0.84 b	$0.19 \pm 0.02a$
Agroforestry system	4.98 ± 0.30 b	2.07 ± 0.30 b	$0.27 \pm 0.02c$	$13.50 \pm 1.47b$	5.58 ± 1.64 b	$0.25 \pm 0.04a$

Note: The values of pH were average of values from samples taken at 3 depths: $0-10 \, \mathrm{cm}$, $10-20 \, \mathrm{cm}$ and $20-30 \, \mathrm{cm}$, but the values of other chemical properties were only from samples of $0-10 \, \mathrm{cm}$. Different letters following values in the same column indicate significant differences among the values (p-value < 0.05), determined with Fisher's LSD test.

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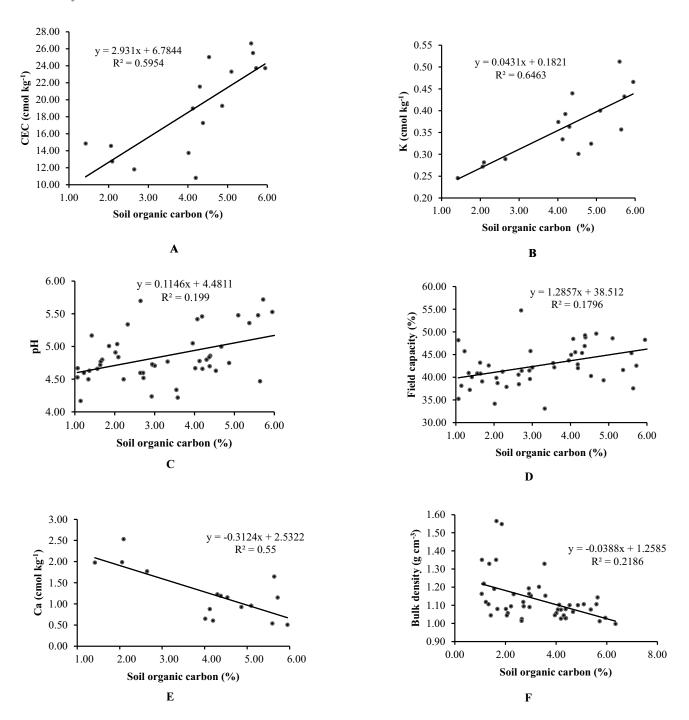


Figure 5 Soil properties having significant correlation with SOC, i.e., A) cation exchange capacity, B) potassium, C) pH, D) field capacity, E) calcium, and F) bulk density.

Soil physical properties and their correlation with SOC Soil bulk density varied with land use types and soil depths (Table 4 and Table 5). The bulk density in the forest was significantly lower than in the agricultural field and agroforestry land, but not significantly different from that in the oil palm plantation. Soil bulk density was the highest in the deep soil layer, while the SOC was the highest in the top layer. Bulk density was negatively correlated with the SOC but the correlation was weak, r = -0.47, p-value < 0.01 (Figure 5F). In the tropical plain of Cameroon, the SOC is

also negatively correlated with bulk density (Kome et al., 2021). The increased SOC may increase soil aggregate stability, reduce compaction, and lower the soil bulk density (McCauley et al., 2017).

The field capacity of the forest soil was significantly higher than in the agricultural field and agroforestry land, but not significantly different from the oil palm plantation. Soil organic carbon had positive correlation with field capacity, but the correlation was weak, r = 0.42, p-value < 0.01 (Figure 5D). In this study, the available water capacity was

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not determined because the laboratory could not do it, so only the field capacity was measured. A positive correlation between available water capacity and the SOC was found in several soils under different vegetation types in Bengkulu Province, Indonesia (Hermawan et al., 2020). In general, soil organic matter was known to increase plant available water by 2 to 3 mm per 10 cm for each 1% increase in soil organic carbon, depending on soil type (Murphy, 2015). Another study found that the relation of SOC and water retention was contradictory (Rawls et al., 2003). At low SOC, an increase in organic matter content increased water retention in sandy soils but decreased it in fine-textured soils. At high SOC, an increase of SOC increased water retention in all soil types.

The soil texture in the study sites was mostly sandy clay loam (Table 6). Soil with loam texture was found only in the top layer of the agroforestry system, clay texture only in the lowest layer of the forest, and clay loam texture only in the lowest layer of the agroforestry land. The forest soil had relatively higher clay content than the other land use types, but the difference was not significantly significant (Table 4). The clay content was not significantly different among soil depths, but there was a tendency that clay increased with the increasing depth (Table 5). Clay content was not correlated with SOC.

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Conclusion

Forest had the highest SOC among the land use types, so if forest is converted to other land uses, the SOC decreases. Higher litter biomass in the forest floor was presumably the cause of higher SOC because litter biomass was positively correlated with SOC. The agricultural land, which was left fallow and occasionally received manure, ranked second in SOC and litter biomass.

Table 4 Soil physical properties in different land use types in the study sites

Land use types	Physical properties			
	Bulk density (g cm ⁻³)	Field capacity (%)	Clay (%)	
Forest	$1.05 \pm 0.03a$	$46.54 \pm 0.03a$	$36.60 \pm 9.40a$	
Agricultural field	$1.17 \pm 0.09b$	$41.32 \pm 0.03b$	$30.51 \pm 14.92a$	
Oil palm plantation	$1.09 \pm 0.05a$	$44.53 \pm 0.04a$	$32.35 \pm 11.31a$	
Agroforestry system	1.20 ± 0.20 b	$38.60 \pm 0.03b$	$24.03 \pm 5.83a$	

Note: Numbers are mean and standard deviation. The values were the average of values from samples taken at 3 depths: $0-10 \,\mathrm{cm}$, $10-20 \,\mathrm{cm}$, and $20-30 \,\mathrm{cm}$. Different letters following values in the same column indicate significant differences among the values (p-value < 0.05), determined with Fisher's LSD test.

Table 5 Soil pH and physical properties at different depths in the study sites

Soil depth (cm)	pН	Bulk density (g cm ⁻³)	Clay (%)	Field capacity (%)
0-10	$5.02 \pm 0.42a$	$1.07 \pm 0.04a$	$26.24 \pm 9.11a$	$42.89 \pm 0.04a$
10–20	$4.86 \pm 0.41b$	$1.11 \pm 0.06a$	$30.81 \pm 10.95a$	$42.30 \pm 0.06a$
20-30	$4.68 \pm 0.24b$	$1.22 \pm 0.17b$	$35.57 \pm 12.75a$	$42.90 \pm 0.04a$

Note: Numbers are mean and standard deviation. Different letters following values in the same column indicate significant differences among the values (p-value < 0.05), determined with Fisher's LSD test.

Table 6 Soil texture in different land use types and depths in the study sites

I and was tomas	Sand	Clay	Silt	Toutum alone
Land use types -	0–10 cm (%)			Texture class
Forest	51.51	31.93	16.56	Sandy clay loam
Agricultural field	48.00	26.53	25.47	Sandy clay loam
Oil palm plantation	52.68	27.54	19.78	Sandy clay loam
Agroforestry system _	44.06	18.96	36.98	Loam
_		10-20 cm (%)		
Forest	49.68	33.67	16.65	Sandy clay loam
Agricultural field	48.76	29.54	21.70	Sandy clay loam
Oil palm plantation	45.03	36.55	18.42	Sandy clay
Agroforestry system _	40.20	23.48	36.31	Sandy clay loam
_				
Forest	39.68	44.21	16.11	Clay
Agricultural field	46.46	35.47	18.07	Sandy clay
Oil palm plantation	48.64	32.97	18.39	Sandy clay loam
Agroforestry system	35.45	29.65	34.9	Clay loam

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Recommendation

Maintaining high litter biomass on the ground is essential to reduce SOC loss when a forest is converted to agricultural land. Therefore, crop residues should be returned, or manure should be added to the soil in the agricultural land..

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